Journal of Tropical Ecology

www.cambridge.org/tro

Interactions between ants and non-myrmecochorous diaspores in a West African montane landscape

Jennifer A. Agaldo[1,*], Alexander V. Christianini[2] and Hazel M. Chapman[1]

[1]School of Biological Sciences, University of Canterbury, 20 Kirkwood Avenue, Upper Riccarton, Christchurch 8041, New Zealand and [2]Universidade Federal de Sao Carlos: Sorocaba, Washington Luis Highway 235km-SP-310, Brazil

Research Article

Cite this article: Agaldo JA, Christianini AV, and Chapman HM. Interactions between ants and non-myrmecochorous diaspores in a West African montane landscape. *Journal of Tropical Ecology*. https://doi.org/10.1017/S0266467420000231

Received: 23 June 2019
Revised: 17 September 2020
Accepted: 16 December 2020

Keywords:
Germination; montane forest; seed cleaning; seed dispersal; seed predation

Author for correspondence:
*Jennifer A. Agaldo,
Email: arubemi@yahoo.com

Abstract

Myrmecochory, the dispersal of seeds with lipid-rich appendages by ants, is a significant ant–plant interaction. Less well understood is the potential for ant dispersal of non-myrmecochorous seeds. Here we investigate ant–diaspore interactions in a West African montane habitat. We combine observation with depot experiments to determine ant species that move diaspores and distance moved across a forest-edge-grassland gradient. We recorded seed cleaning by ants using a bird/mammal dispersed *Paullinia pinnata* to determine whether seed cleaning improved plant fitness. We found that two out of a total of 17 ant species (*Pheidole* sp. 1 and *Myrmicaria opaciventris*) interacted with 10 species of non-myrmecochorous diaspores across nine plant families. Diaspores were from large canopy trees, understorey trees and vines. Both ant species interacted with small (≤0.24 g) and large (≥0.24 g) diaspores. Ants individually moved small diaspores up to 1.2 m and worked together to clean larger ones. Our experiments with *P. pinnata* showed that ants removed the pulp of 70% of fruit over 5 days. Cleaned seeds germinated significantly faster and produced seedlings with significantly longer shoot length and higher fresh weight than seedlings from intact seeds. Together our results suggest that ant dispersal may be less significant than seed cleaning in Afromontane forests. However, given the decline in vertebrate frugivores across Africa, a small dispersal advantage may become increasingly important to plant fitness.

Introduction

Seed dispersal is an essential process in plant ecology (Nathan & Muller-Landau 2000, Wang & Smith 2002), influencing fitness at the level of the individual and species (Nathan & Muller-Landau 2000). At the community level, dispersal underpins patterns of spatial recruitment (Howe & Miriti 2000, Seidler & Plotkin 2006), strongly influencing the trajectory of species composition across generations. Factors affecting seed dispersal are numerous, including diaspore (i.e. unit of dispersal – seed or fruit) type, size and seed disperser guild (Tamme *et al.* 2014, Vittoz & Engler 2007). In tropical forests the majority of woody plant species have fleshy fruits adapted for dispersal by a range of generalist frugivores (Van Schaik *et al.* 1993), thus dispersal is often non-precise, with seeds being scattered and landing in sites not necessarily optimal for germination and establishment (Wenny 2001). Therefore, despite trees often producing fruit crops far in excess of what frugivores can consume (an adaptation to attract frugivores), the majority of species suffer severe dispersal limitation (Hubbell *et al.* 1999, Svenning *et al.* 2006). Undispersed fruits fall to the ground where they may suffer predation by animals, or death through intraspecific competition and/or pathogens (Janzen 1970). However, increasing evidence suggests that seeds dispersed into non-optimal habitats or those fallen under the parent tree, are not necessarily destined to such fates. Animals can collect seeds from the ground, move them to different locations and store them for later consumption – e.g. scatterhoarding in caches by rodents (Gómez *et al.* 2019) or taken into nests by ants to feed colony members (Pizo & Oliveira 2001). When such seeds are lost or forgotten, they are often in prime locations for germination and establishment (Camargo *et al.* 2016, Gómez *et al.* 2019). Following the pioneering work of Roberts & Heithaus (1986) on the dispersal of fig seeds by ants, more recent work in the Neotropics has demonstrated the potential for ants to disperse small (<1 g) seeds from non-myrmecochorous diaspores (Passos & Oliveira 2002, Pizo & Oliveira 2001) either directly from fallen fruits or by moving seeds first dispersed by birds, into 'safe' ant nest sites (Christianini & Oliveira 2010, Passos & Oliveira 2002, 2004). In this way, ants can alter seed shadows, with potential benefits for seed germination and seedling establishment (Camargo *et al.* 2016, Christianini & Oliveira 2010). Even if not dispersing seeds, ants may clean seed of fruit pulp where they find them on the ground, which can benefit seed germination (Gallegos *et al.* 2014, Oliveira *et al.* 1995). These benefits may vary across habitats such as forests and savannas (Camargo *et al.* 2019), and across habitat edge boundaries (Christianini & Oliveira 2013,

© The Author(s), 2021. Published by Cambridge University Press

CAMBRIDGE UNIVERSITY PRESS

Downloaded from https://www.cambridge.org/core. University of Canterbury Library, on 29 Apr 2021 at 21:09:13, subject to the Cambridge Core terms of use, available at https://www.cambridge.org/core/terms. https://doi.org/10.1017/S0266467420000231

Gallegos *et al.* 2014). While seed dispersal by large ants is often reduced in disturbed sites (Leal *et al.* 2014), small ants may still have a positive impact on regeneration of some species in sites disturbed or under restoration (Gallegos *et al.* 2014, Zwiener *et al.* 2012). The participation of ants in interactions with fruits/seeds seems pervasive at community level (Christianini *et al.* 2012), and benefits are often asymmetric between interaction partners (Giladi 2006, Warren & Giladi 2014).

While lowland tropical forests are known to harbour high ant diversity and abundance (Guénard *et al.* 2012) with many species interacting with fruits and seeds (Holldobler & Wilson 1990), so that their role in seed dispersal is not surprising, the extent to which ants interact with seeds in tropical montane forests, where ants are less diverse and abundant (Bruhl *et al.* 1999, Dunn *et al.* 2009, Sabu *et al.* 2011), is unclear. However, the fact that a higher relative frequency of tropical montane forest species produce small, bird-dispersed fleshy fruits than lowland tropical forest species (Chapman *et al.* 2016) suggests that ants may play a role in seed dispersal after seeds reach the ground and may contribute to plant fitness. To answer this question, here we investigate for the first time, the nature of ant–diaspore interactions in a West African montane landscape. As this is the first study of its kind in this locale, we first identified all ant species we found interacting with diaspores (Agaldo 2017). Then, based on findings from previous studies elsewhere, we tested four hypotheses:

(i) Ants will interact with non-myrmecochorous diaspore species (Christianini *et al.* 2012, Kaspari 1993, Levey & Byrne 1993, Pizo & Oliveira 1998) by moving and cleaning.
(ii) Ants will clean the flesh of some diaspore species which will enhance seed germination and seedling establishment (Christianini *et al.* 2007, Leal *et al.* 2007, Oliveira *et al.* 1995).
(iii) Different communities of ants associated with grasslands, edge habitat and core forest will lead to different dispersal distances in the three habitats (Leal *et al.* 2014, Zelikova & Breed 2008).
(iv) A relatively large number of ant species will move a relatively small number of diaspore species (Christianini *et al.* 2012, Passos & Oliveira 2003, Zwiener *et al.* 2012).

Methods

Study area

Our study area, Ngel Nyaki Forest Reserve (NNFR) (7.08°N, 11.12°E; Figure 1), is situated in the mountains of Taraba State, Nigeria between 1400–1600 m in elevation (Barnes & Chapman 2014, Chapman 2008), with a distinct dry (November–March) and wet season (April–October) (Matthesius *et al.* 2011). The 46 km^2 reserve is a mosaic of grassland and forest; there is a 5.2 km^2 patch of submontane escarpment forest and small, degraded riparian forest fragments embedded within a matrix of heavily overgrazed *Sporobolus pyramidalis* grassland (Barnes & Chapman 2014, Chapman 2008). The forest has a sharp boundary with the grassland, but when protected from grazing and fire the grasslands become more diverse with increased grass species and woody shrubs (Barnes & Chapman 2014). The escarpment forest has a diverse plant and animal community (Beck & Chapman 2008, Chapman *et al.* 2001). There are at least 17 ant morpho species in the area, dominated by an unidentified *Pheidole* sp. and *Myrmicaria opaciventris* (Agaldo unpubl. data).

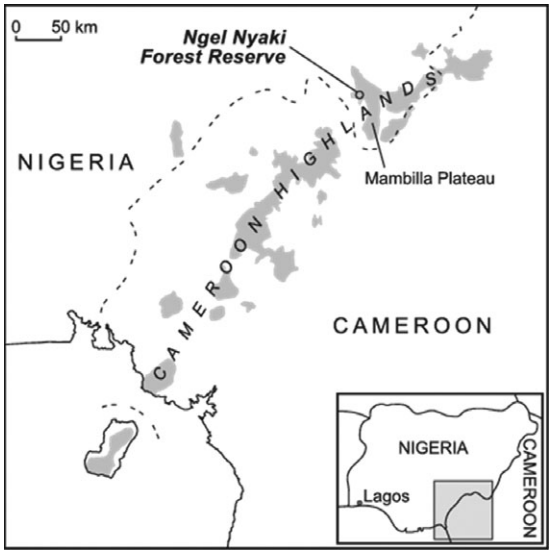

Figure 1. Map of Ngel Nyaki Forest Reserve showing location of study sites. Adapted from Google Maps.

Study design

Because our main aim was to investigate the most common ant–plant interactions, rather than create an exhaustive list of ant species, we used a combination of different sampling methods including active searches (reconnaissance surveys) and baiting methods (removal experiments) as recommended by Agosti *et al.* (2000). We based our investigations in three sites within NNFR (Figure 1). Sites, at least 0.5 km apart, comprised three habitats: core escarpment forest (hereafter referred to as forest), forest edge (edge) and grassland. Following Gallegos *et al.* (2014), at each site we ran two transects separated by a distance of 20 m, each 320 m long, from the core forest, across forest edge and 160 m into the grassland. Observations and experiments were conducted between June 2015 and September 2017.

Reconnaissance surveys

Our reconnaissance surveys involved walking each transect once a month between June and September 2015. Walks were taken between 0730 h – 1300 h, the hours we found ants to be most active, and involved moving at a slow pace (~1 km/h) while searching on the ground ~2 metres on either side of the transect for ant–diaspore interactions. For every interaction observed we recorded interaction type (cleaning/removing tissue on the surface or moving), ant species and plant species (Passos & Oliveira 2003). We collected samples of each ant species and stored them in 70% ethanol for later identification using the taxonomic keys of Bolton (1994). Ant identifications were checked and confirmed by an expert ant taxonomist at Universidade Federal do Paraná, Brazil.

Removal experiments

To determine the types of diaspore moved and to what distances, we carried out removal experiments following the methods and sampling intensity of Christianini *et al.* (2007, 2012). First, we collected fruits from 13 woody plant species with small fruit weighing ≤1 g, a weight realistically moved by ants (Pizo & Oliveira 2001). The number of diaspore species we used was

Downloaded from https://www.cambridge.org/core. University of Canterbury Library, on 29 Apr 2021 at 21:09:13, subject to the Cambridge Core terms of use, available at https://www.cambridge.org/core/terms. https://doi.org/10.1017/S0266467420000231

Table 1. Plant species used for diaspore removal experiment in Ngel Nyaki Forest Reserve.

Plant species	Family	Mean number removed (n)	Ant removal	Mean distance moved (cm) (min–max)
*Allophylus africanus***	Sapindaceae	0 (156)	No	
*Bridelia speciose***	Euphorbiaceae	0.08 (156)	Yes	8.5 (3–16.5)
*Celtis gomphophylla**	Cannabaceae	0.04 (251)	Yes	29.5 (3.5–59)
*Diospyros cameroonensis***	Ebenaceae	0 (156)	No	–
*Harungana madagascariensis****	Hyparicaceae	0.13 (155)	Yes	19.5 (3–80)
*Macaranga monandara***	Euphorbiaceae	0 (156)	No	–
*Maesa lanceolata****	Myrsinaceae	0 (156)	No	–
*Polyscias fulva***	Araliaceae	0.08 (368)	Yes	34.74 (17.2–58.5)
*Psorospermum aurantiacum****	Guttifereae	0 (156)	No	–
*Psychotria umbellata****	Rubiaceae	0 (156)	No	–
*Rauvolfia vomitoria***	Apocynaceae	0 (156)	No	–
*Trema orientalis***	Cannabaceae	0.40 (307)	Yes	29.34 (2–120.3)
*Zanthoxylum leprieurii***	Rutaceae	0.78 (78)	Yes	25.9 (2.6–65.2)

*Core forest tree; **Forest edge/grassland pioneer tree; ***Grassland shrub/tree.

limited by the availability of species fruiting at the time of the experiments (Table 1). For each species used in the diaspore removal experiments, we collected fruits from five individual parent trees and removed the seeds from a random sample of these fruits (for seeds used in the experiment), washed the seeds to remove any fruit pulp or fleshy appendage and sun-dried them for 1 h (Roberts & Heithaus 1986). This was done to mimic seed conditions after regurgitation by birds.

In each experiment we established 13 depots on the ground, 1 m away from the transects described above. Depots comprised a piece of white filter paper (Pizo & Oliveira 2001) on which we placed two fruits and two seeds of the same species. We placed the depots at 5 m, 10 m, 20 m, 40 m, 80 m and 160 m from the forest edge in both directions (i.e. into grassland and forest). Two of us, positioned at least 0.5 m on either side of each depot, recorded every ant–diaspore interaction over a 15 min period. We recorded interactions as: (1) removed diaspore to the nest, (2) consumed diaspore on the spot, (3) no removal/interaction.

We followed all ants that removed diaspores from the point of interaction at the depot to where they deposited or dropped the diaspore and measured the distance to the nearest cm. Experiments were again carried out between 0730 h and 1300 h (Agaldo J. pers obs).

Seed cleaning

Paullinia pinnata (Sapindaceae), widespread across Africa (Zamble *et al.* 2006), is a common woody vine of NNFR forest edge (Chapman *et al.* 2001). Its fruits, 25 mm (± 3) long and (20 ± 2 mm) wide are red when ripe, containing 2–3 seeds 13.5 (± 2) mm long and 7 mm (± 2) wide (NMFP fruit trait data), covered in a dense white aril (Weckerle & Rutishauser 2005). When the fruits ripen, the seeds are dispersed by primates and birds (Mascaro *et al.* 2004).

We knew from initial observations that the *P. pinnata* aril attracted two common ant species, *Pheidole* sp. 1 and *Mymicaria opaciventris*. To determine the rate at which these ants remove the aril and cleaned *P. pinnata* seeds, we carried out a 5-day experiment, again, based on seed depots. The depots comprised white filter paper, but this time each depot was protected with a mesh cage (0.5 × 0.5 cm mesh) placed over the filter paper and pinned to the ground to exclude large invertebrates and vertebrates. We placed one seed of *P. pinnata* in each depot. Because *P. pinnata* is a forest edge species, we placed depots in edge habitat. Our first depot was placed under the outermost canopy tree (0 m), with four more depots at 10 m intervals to 40 m into the forest. We added the seeds to the depots at 0730 h and observed the ant seed cleaning activity on days 1, 2, 3 and 5 at 0730 h. During each observation period, we recorded ant species and number of ants cleaning seeds. We visually assessed the amount of seed appendage removed ('cleaned') by ants and grouped the seeds into one of three categories (Pizo & Oliveira 1998): (i) Intact seeds – seeds to which ants paid no attention and had 0% of its aril missing; (ii) still being cleaned – seeds that had between 1–75% of aril missing; and (iii) cleaned – seeds that had between 76–100% of the aril removed by ants.

The effect of ant cleaning

To determine whether seed 'cleaning' affects the germination rate and/or seedling biomass of *P. pinnata*, we designed an experiment to compare: (i) days to emergence of the radicle, (ii) seedling growth rate, (iii) seedling total fresh biomass between intact and cleaned seeds.

During the wet season (July and August 2016), we collected 36 seeds from below five *P. pinnata* parent plants, making sure all the seeds had been dispersed through ballistic expulsion and were free of any observable insect attack. From these seeds we created two treatments: (i) intact seeds, as they were when collected from the forest floor and (ii) thoroughly cleaned seed, prepared by manually removing all of the aril. We sowed the seeds 5 cm deep into separate 6-litre, black polythene 'pots' filled with regular commercial compost. Pots were placed in the plant nursery in a blocked design at the NMFP and checked for germination and growth (measured with a mm tape measure) every 10 days over a period of 40 days. Sixty days after planting, we counted the number of leaves on germinated seedlings in each treatment and the length of each shoot. Following this, we harvested the shoots and roots, measured root lengths and weighed the shoots and roots separately.

Downloaded from https://www.cambridge.org/core. University of Canterbury Library, on 29 Apr 2021 at 21:09:13, subject to the Cambridge Core terms of use, available at https://www.cambridge.org/core/terms. https://doi.org/10.1017/S0266467420000231

Table 2. Opportunistic ant–diaspore interactions recorded during reconnaissance survey in Ngel Nyaki Forest Reserve.

Plant species	Ant species	Diaspore type/portion	Interaction type
Psorospermum aurantiacum	*Pheidole* sp. 1 *M. opaciventris*	Fruit Fruit pulp	Moved Cleaned on the spot
Paullinia pinnata	*Pheidole* sp. 1 *M. opaciventris*	Seed aril Seed aril	Cleaned on the spot Cleaned on the spot
Syzygium macrocarpa	*Pheidole* sp. 1	Fruit pulp	Cleaned on the spot
Trema orientalis	*Pheidole* sp. 1 *M. opaciventris*	Fruit and seed Fruit and seed	Moved Moved
Chionanthus africanus	*Pheidole* sp. 1	Fruit pulp	Cleaned on the spot
Zanthoxylum leprieurii	*M. opaciventris*	Fruit	Moved

Data analysis

We analysed all data using R version 3.2.4 with *lme4* and *car* packages. For all analysis, we considered a P-value less than 0.05 as significant.

Spatial and temporal variation in ant diaspore removal

We used a generalized linear mixed-effect model (GLMM) with a binomial error distribution to assess spatial and temporal variations in ant diaspore removal. The response variable was a combination of diaspore 'removed' and 'not removed', while independent variables were fixed effects which included 'habitat', 'season – wet and dry', 'ant species', 'plant species' and their two-way interactions. The random effects included in the models were 'transect ID', 'site' and 'depot ID'.

Effect of ant species and plant species on diaspore removal distance

We used a GLMM with Poisson error distribution to assess how the different ant and plant species affected diaspore removal distance. The response variable was the distance moved and the independent variables were the fixed effects which included 'ant species', 'plant species' and their interactions. Random effects were 'transect ID', 'depot' and 'observation ID'. We compared all models using one-way ANOVA and removed factors with the least significant effects. Akaike Information Criteria (AIC) were used to choose the best-fit model (Crawley 2002, Symonds & Moussalli 2011). When models were not significantly different, we chose the one with the fewest factors as the best-fit model. We used the *blmeco* package in R for model validation by checking for overdispersion. An observational-level random effect was included in models to account for overdispersion (Harrison 2014).

Seed cleaning rate of Paullinia pinnata

We summarized our census based on the three seed categories (intact seeds, still being cleaned, or cleaned) and used an independent t-test to test the difference in the numbers of *Pheidole.* sp. 1 and *M. opaciventris* cleaning diaspores of *P. pinnata*.

Effect of seed cleaning on Paullinia pinnata *germination and seedling growth*

We used the G-test of independence to determine if there was a significant difference in the proportion of cleaned versus intact diaspores that germinated. Shoot length, root length, number of leaves and fresh weight of seedlings from intact and cleaned seeds were compared using the unpaired Wilcoxon–Mann–Whitney test.

Table 3. Results of the generalized linear mixed model with binomial distribution showing the effect of habitat, season, ant species and plant species on diaspore removal. Significance is denoted by an asterisk at $P \leq 0.05$.

Sources of variation	χ^2	df	P
Habitat	3.2	2	0.19
Season	4.59	1	<0.05*
Ant species	51.2	2	<0.0001*
Plant species	30.1	4	<0.0001*

Results

Ant–diaspore interactions

Our 16 h of walking transects yielded 18 ant–diaspore encounters. Two ant species, *Myrmicaria opaciventris* and *Pheidole* sp. 1 (both Myrmicinae), interacted with diaspores from six plant species (Table 2). Interactions included moving fruits or seeds, cleaning fleshy pulp from fruits and foraging on seed appendages such as the arils of *P. pinnata.*

Spatial and temporal variation in diaspore removal

Nineteen per cent (250) of the 1315 diaspores used in the removal experiment were removed by ants. While habitat (grassland, edge or forest) did not affect diaspore removal, season and plant species had significant effects (Table 3). Ants removed more diaspores of *Zanthoxylum leprieurii* and *Trema orientalis* than those of *Bridelia speciosa*, *Celtis gomphophylla*, *Polyscias fulva* and *Harungana madagascariensis* (Table 1). The mean frequency of diaspore removal was higher during the wet season (0.29 ± 0.02, mean ± SE) than the dry season (0.06 ± 0.01).

Effect of ant and plant species on removal distance

Seventy per cent of the diaspores we observed being moved by ants were 'lost' beneath leaf litter and we were unable to follow them further. However, we were able to follow the remaining 30% (75 interactions) all the way into ant nests. Of these, removal distance did not differ across habitats or between ant species but was highly variable within and among plant species. Dispersal distances ranged from 2 cm to 1.24 m with a mean of 28.6 cm, and median of 24 cm (Table 1).

Downloaded from https://www.cambridge.org/core. University of Canterbury Library, on 29 Apr 2021 at 21:09:13, subject to the Cambridge Core terms of use, available at https://www.cambridge.org/core/terms. https://doi.org/10.1017/S0266467420000231

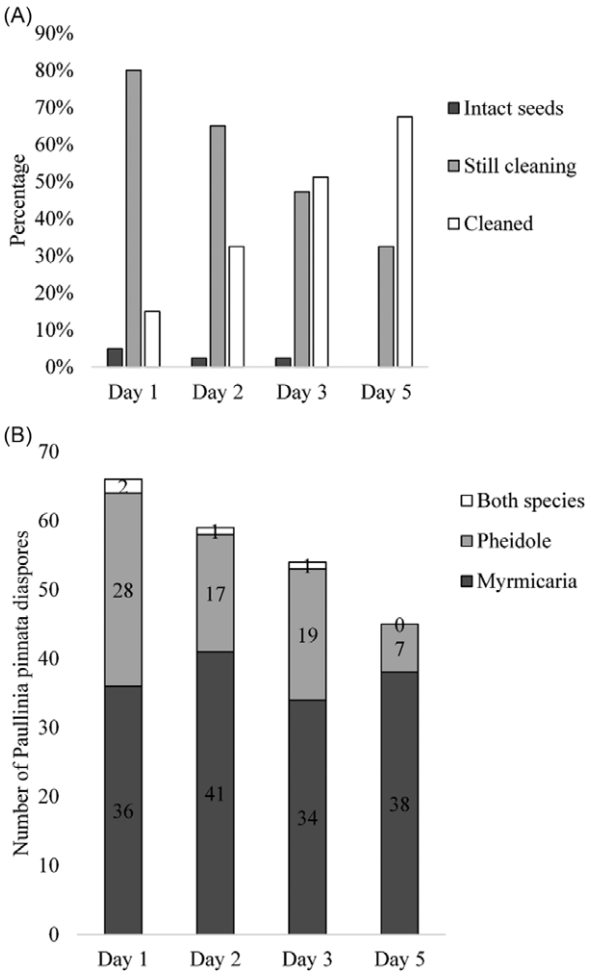

Figure 2. (A) Seed cleaning rates of *Paullinia pinnata* by ants, over a 5-day period (number of seeds = 80). (B) Ant species attendance to diaspores of *Paullinia pinnata* over 5 days. Labels on bars indicate number of depots where *Pheidole* sp. 1 and *Myrmicaria* attended seeds either separately or together.

Pheidole sp. 1 was responsible for the majority of all diaspore removals (74%) and *M. opaciventris* 26%. Of the diaspores moved by *Pheidole* sp. 1, we observed 36% being moved into ant nests with the remaining 64% being lost from view under leaf litter. Only 12% of diaspores moved by *M. opaciventris* were followed until the nests, while 88% disappeared into leaf litter.

Seed cleaning rate of P. pinnata

The proportion of seeds cleaned by ants versus those left alone or only partly cleaned after days 1, 2, 3 and 5 of the experiment, are summarized in Figure 2. We found a significant difference between the two ant species (*Pheidole* sp. 1 and *M. opaciventris*) in the number of ants cleaning seeds ($t = 4.27$, $df = 6$, $P < 0.005$) with a mean of 37.35 ± 1.49 (mean ± SE) for *M. opaciventris* and 17.75 ± 4.31 for *Pheidole* sp. 1 (Figure 3). We never observed the two ant species visiting diaspores at the same time.

Effect of cleaning on the growth of P. pinnata *seedlings*

Overall, our results suggest that cleaning has a positive effect on seed germination and seedling growth. Shoots from cleaned seeds were on average longer (6.17 ± 0.28 cm, mean ± SE) than intact seeds (4.71 ± 0.26 cm) ($W = 270$, $P < 0.001$) (Figure 3). Seedlings from cleaned seeds produced more leaves ($W = 217.5$, $P < 0.05$) and had a higher mean fresh weight ($W = 157.5$, $P = 0.02$) than those from intact seeds. Mean fresh weight of seedlings from cleaned seeds was 1.21 ±0.14 g against 0.78 ± 0.13 g from intact seeds. In contrast, the root lengths of seedlings from intact (22 ± 1.66 cm) and cleaned seeds (24.3 ± 1.22 cm) did not differ significantly ($t = 0.89$, $df = 25.3$, $P = 0.37$) (Figure 3).

Discussion

Ant–diaspore interactions

Based on a combination of active searches and depot experiments across three habitats we found only two ant species interacting with 10 non-myrmecochorous diaspore species in ways which may enhance plant fitness: a novel finding with important potential implications for plant regeneration in African forests. Out of 17 species of ground-dwelling ants recorded in NNFR (Agaldo 2018), we observed the small-bodied species *Pheidole* sp. 1 and *Myrmicaria opaciventris* (both Myrmicinae), interacting with diaspores of small seeded woody plant species from across the forest. These findings support our first hypothesis based on reports from the Neotropics, that ants may interact in positive ways with non-myrmecochorous diaspores and potentially influence seed fate (Christianini *et al.* 2012, Kaspari 1993, Levey & Byrne 1993, Passos & Oliveira 2002, Pizo & Oliveira 1998).

Different ant species can have very different interactions with diaspores, with major implications for plant distribution (Magalhães *et al.* 2018). While both ant species we found interacting with diaspores in our study area are small bodied and short ranging, characteristics of the Myrmicinae (Andersen 1995), some of their behaviours differed in ways which could potentially influence plant fitness. Arguably most important is foraging guild – both ant genera include generalist omnivores (Hölldobler & Wilson 1990), but their diets differ. *Pheidole* species are granivorous (Levey & Byrne 1993, O'Dowd & Gill 1984, Pirk *et al.* 2009) and therefore seed predators. In contrast, *M. opaciventris* feeds on insects and honeydew (Kenne & Dejean 1997), with our study being the first to report *M. opaciventris* interacting with plant diaspores. Another difference is in propensity to move, versus clean, diaspores. *Pheidole* sp. 1 moved more diaspores from our depots (75%), while *M. opaciventris* did more seed cleaning on the spot (67%). Moreover *Pheidole* sp. 1 was responsible for all the distances moved > 1 m and was the only species we observed recruiting to help move seeds. Other species in the genus *Pheidole* are known to disperse seeds short (< 50 cm) distances (Andersen 1995, Thomson *et al.* 2016, Zelikova & Breed 2008), but *Pheidole fallax* in dry Costa Rican forests recruits to move seeds greater distances, up to 3 m (Zelikova & Breed 2008).

Seed dispersal

We found 19% of diaspores, comprising six of the 13 diaspore species included in the experiment, were moved from depots within 15 min. Fast removal of plant diaspores from the soil surface may decrease seed exposure to predators and enhance plant fitness (Christian 2001, Thomson *et al.* 2016). However, given that the small fruits of *Psorospermum aurantiacum* were not moved in the experiments, but were seen being moved during the reconnaissance survey, suggests that given greater sampling intensity, more

Downloaded from https://www.cambridge.org/core. University of Canterbury Library, on 29 Apr 2021 at 21:09:13, subject to the Cambridge Core terms of use, available at https://www.cambridge.org/core/terms. https://doi.org/10.1017/S0266467420000231

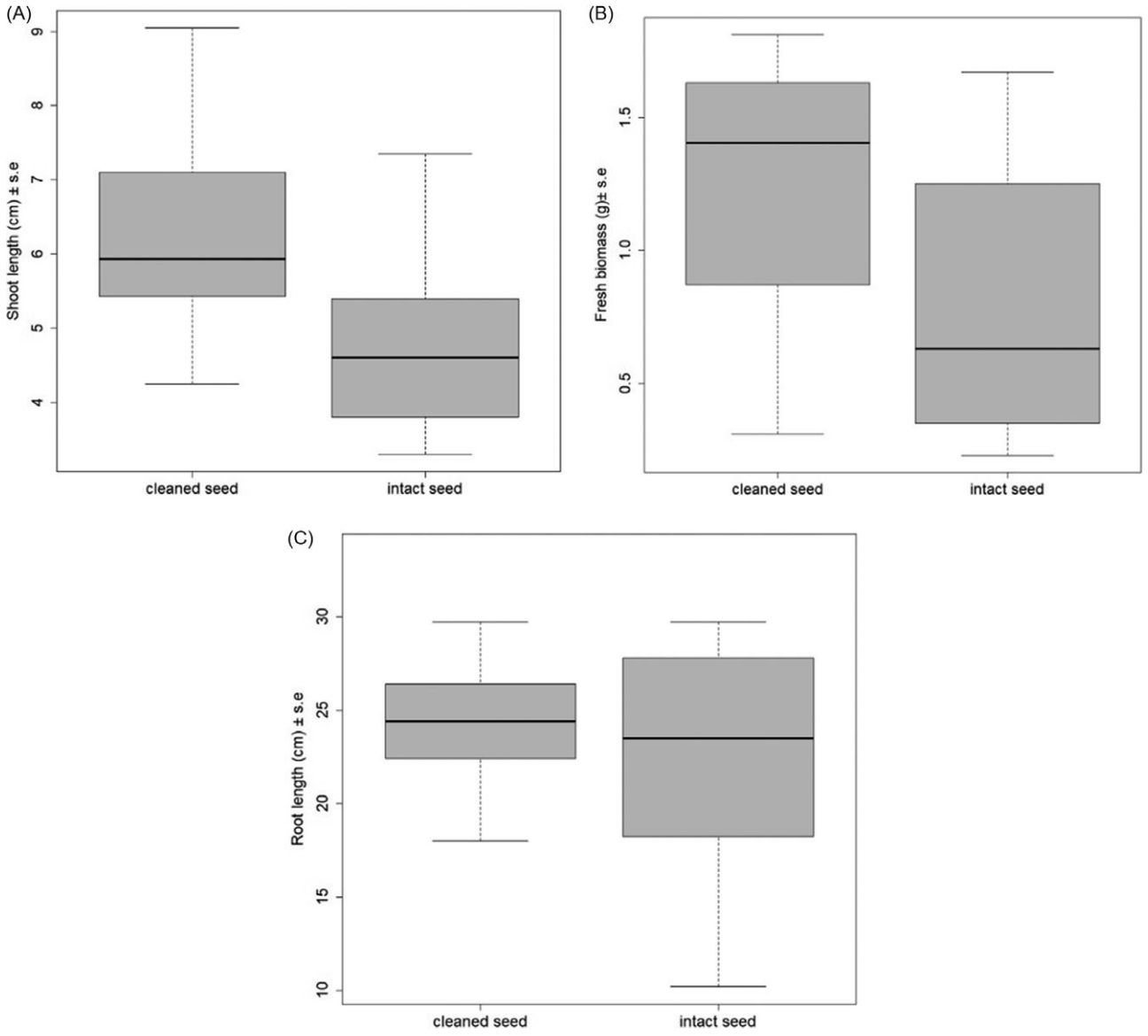

Figure 3. (A) shoot length (cm), (B) wet weight (g) and (C) root length (cm) of seedlings from cleaned and intact seeds of *P. pinnata* (n = 18 for cleaned seeds, n = 18 for intact seeds). Boxes indicate median and quartiles, with bar limits depicting upper and lower ranges.

of the 13 species presented may be moved. All diaspores which were moved weighed ≤0.24 g, and mostly comprised fruits and seeds of forest edge/grassland pioneer species (e.g. *Polyscias fulva* and *Trema orientalis*) with only one core forest species, the large canopy tree (*Celtis gomphophylla*) (Table 1).

The maximum distance we observed a diaspore being moved was 124 cm, with the mean distance being only 24 cm. This is considerably less than the 1.99 m estimated by Gómez & Espadaler (2013) as the global mean removal distance of myrmecochorous diaspores by ants and the mean of 5 m recorded for non-myrmecochorous diaspores in Neotropical savanna (Christianini *et al.* 2007). It is also less than the 3 m recorded for *P. fallax* in Costa Rica forest (Zelikova & Breed 2008). Such relatively short distances in NNFR are not unsurprising, since small ants tend to move diaspores shorter distances than large-bodied ants (Andersen 1988, Gómez & Espadaler 2013, Ness *et al.* 2004). Both density and distribution of ant nests affect foraging (Andersen 1988). While we did not measure nest density during this study, the fact that we observed ants moving diaspores to nest locations not far from experimental depots does suggest that nests may be abundant. Abundant food sources can also decrease the foraging distances of ants by reducing home ranges (Bernstein 1975). However, as seed production tends to be limited in montane relative to lowland sites (Chapman *et al.* 2016) this argument needs further testing. Whatever the reason, and even with the possibility that we missed rare longer distance events, dispersal distances are short in this montane habitat, which may make it unlikely that ants help seeds escape intra-specific, negative density-dependent mortality (Connell 1971, Janzen 1970) and gain associated benefits (Andersen 1988, Cain *et al.* 2000, Gallegos *et al.* 2014). It is worth noting that in Afromontane forests, tree species tend to show more local clumping than in lowland tropical forests (Abiem *et al.* 2020, Matthesius 2006) and Janzen–Connell effects may not be as pervasive (Matthesius *et al.* 2011). With short-distance movement of

Downloaded from https://www.cambridge.org/core. University of Canterbury Library, on 29 Apr 2021 at 21:09:13, subject to the Cambridge Core terms of use, available at https://www.cambridge.org/core/terms. https://doi.org/10.1017/S0266467420000231

Table 4. Results of the generalized linear mixed model with Poisson distribution showing the effect of habitat, ant species and plant species on diaspore removal distance. Significance is denoted by an asterisk at P ≤ 0.05.

Source of variation	χ^2	df	P
Habitat	0.02	2	0.98
Ant species	0.08	2	0.76
Plant species	22.5	5	< 0.001*

diaspores into 'safe sites' such as ant nests, the probability of germination and establishment may increase (Farji-Brener & Werenkraut 2017). Thus, even short distance movement may be important, as ant nests or even leaf litter can afford seeds protection from desiccation (Giladi 2006) and predators (Gorb & Gorb 2003, Passos & Oliveira, 2004). Rapid ant diaspore removal may reduce diaspore loss to vertebrate predators (Thomson *et al.* 2016). We found some evidence of this from recording the fate of *Trema orientalis* seeds moved by ants from the depots and deposited into their nests; after 4 months *T. orientalis* seedlings were observed growing out of seven of 21 nests (J.A. Agaldo, pers. obs. and unpubl. data). Levey & Byrne (1993) show that *Pheidole* sp. can play a dual role as both seed disperser and predator. From our study, although *Pheidole* was observed moving diaspores, we were not able to determine if it was acting as a predator or a disperser. Seedlings of *T. orientalis* that we observed growing in nests were seen in nests of *M. opaciventris.* Of course, the fate of these seedlings is what matters, and we have no data on the proportion of these seedlings recruited into the next generation, which is a critical question. Overall, our study demonstrates that ants do move a range of small diaspores over short distances and into safe sites. But to what extent ants in this montane ecosystem affect plant fitness remains unclear, and long-term experimentation will be needed to answer this question.

Seed cleaning

From our reconnaissance survey we found that both *Pheidole* sp. 1 and *M. opaciventris* cleaned seeds. Our hypothesis that ants will clean seeds and thus enhance germination and seedling growth rate was supported in our experiment with *Paullinia pinnata.* We found cleaned *P. pinnata* seeds germinated at a faster rate than uncleaned diaspores and their seedlings were larger, gaining more fresh weight. Larger seedlings tend to show higher resistance to environmental stressors such as herbivory, trampling and desiccation than small seedlings, which is linked to higher early survival (Moles & Westoby 2004, Verdú & Traveset, 2005) and, of relevance to NNFR, may be important in increasing resistance to seasonal variability in moisture (Orrock & Christopher 2010). Faster germinating diaspores are also more likely to avoid seed predation because of the limited period of exposure to granivores (Bowers & Dooley 1993, Christianini *et al.* 2007). In addition to *P. pinnata* we observed ants cleaning diaspores of *Psorospermum aurantiacum*, *Syzygium macrocarpa* and *Chionanthus africanus.* It will be important to determine the extent this cleaning aids in seedling recruitment of these species as well.

Different habitats

We found no support for our hypothesis that ant–diaspore interactions would vary among habitats (grassland, edge and forest) (Table 4). While studies elsewhere have shown that removal rates and dispersal distance tend to be modified with disturbance (Christianini & Oliveira 2013, Gallegos *et al.* 2014, Leal *et al.* 2014), we found no variation across habitats despite the fact that both forest edge and grassland, in contrast to core forest, are highly disturbed. This is most likely because assemblage composition strongly impacts dispersal distances (Palfi *et al.* 2017), and in our study the same two small-bodied myrmicinae species are moving seeds across all three habitats. These two ants are the most abundant species in NNFR (J.A. Agaldo, unpubl. data) and it is also possible that they are resilient to disturbances. Indeed, some ant species can persist and interact with diaspores in simplified and disturbed environments (Raimundo *et al.* 2004). *Myrmicaria opaciventris* has been found to occur in and forage in highly altered landscapes such as agricultural fields (Kenne & Dejean 1999).

Asymmetrical interactions

Our findings differ significantly from Neotropical reports in the number of ant vs diaspore species involved in the network, so that our final hypothesis was unsupported. For example, while we found only two ant species (*Pheidole* sp. 1 and *M. opaciventris*) interacting with 10 diaspore species, Passos & Oliveira (2003) reported 41 ant species interacting with diaspores of 56 plant species in the Brazilian lowland Atlantic forest. One explanation for this difference may be the relatively low ant species richness and diversity recorded at Ngel Nyaki compared with neotropical forests, which contain some of the most diverse ant assemblages on Earth (Guénard *et al.* 2012). The smaller number of ant species interacting with fruits may also be due to the pattern of decreasing ant species richness with elevation (Smith *et al.* 2015), since NNFR is at a higher altitude than the sites where other studies were performed. Notwithstanding, our findings support the hypothesis that interactions between ants and diaspores are diffuse and asymmetrical (Warren & Giladi 2014).

Limitations

Given our limited sampling and the opportunistic nature of the interactions, increased observation hours and increasing the number of diaspore species used in depot experiments would inevitably increase the diversity of plant species in the interactions we observed. However, we designed our sampling to capture the most common ant–plant interactions and we are confident we have done this. While there is no comparable study elsewhere from African forests to establish comparisons, a similar study of the number of ant species recorded in interactions with true myrmecochorous seeds of Proteaceae in Fynbos shrublands in South Africa indicated that only two ant species performed the bulk (70%) of ant–seed interactions (Christian 2001). Indeed, it is not uncommon for a small subset of the ant community to dominate the interactions with seeds at a given site (Gove *et al.* 2007, Warren & Giladi 2014).

Conclusion

Our study suggests that ants are unlikely to play a major role in the dispersal of non-myrmecochorous diaspores in Afromontane forests, although the significance of movement into safe sites such as nests needs further investigation. By seed cleaning, ants may enhance germination and seedling growth rates in some species, especially those from the forest edge and grassland. Further investigations are essential to expand our findings, since benefits to plants from ant–diaspore interactions can be subtle but

Downloaded from https://www.cambridge.org/core. University of Canterbury Library, on 29 Apr 2021 at 21:09:13, subject to the Cambridge Core terms of use, available at https://www.cambridge.org/core/terms. https://doi.org/10.1017/S0266467420000231

nevertheless important to plant fitness (Camargo *et al.* 2016). Irrespective, seed shadows and spatial patterns of seedling recruitment within the reserve may be altered through the removal of diaspores by ants (Zelikova 2008), with implications for forest structure.

Based on our findings, we suggest that generalist, opportunistic ant interactions with non-myrmecochorous diaspores will probably be found across the Afrotropics, with more species interacting in lowland than montane forest habitats. It is possible that in Afromontane habitats, diaspore cleaning leading to reduced seed mortality and increased germination may be the more beneficial interaction for non-myrmecochorous diaspores. Given the pervasive reduction in populations of vertebrate frugivores due to habitat loss and hunting that cascades to decreasing seed removal and dispersal distances from parental plants (e.g. Cordeiro & Howe 2003), we believe ants may play an increasing role in seed germination and seedling fate of non-myrmecochorous diaspores in African tropical forests.

Acknowledgements. Retired General TY Danjuma for funding, staff of Nigeria Montane Forest Project for support during fieldwork, Musa Bawuro (RIP) and Thomas, Godson Dangu and Hannah Samuel our field assistants. Dr Shola Olaniyan (Lincoln University New Zealand) and Dr Rodrigo M. Feitosa (Universidade Federal do Paraná, Brazil) for help with ant identifications.

References

Abiem I, Arellano G, Kenfack D and Chapman H (2020) Afromontane forest diversity and the role of grassland-forest transition in tree species distribution. *Diversity* **12**, 30.

Agaldo JA (2017) The role of ants in seed dispersal and regeneration in a degraded West African montane forest-grassland landscape, Ngel Nyaki Forest Reserve, Nigeria Thesis. University of Canterbury.

Agosti D, Majer JD, Alonso, LE and Schultz TR (2000) Standard methods for measuring and monitoring biodiversity. Smithsonian Institution, Washington DC, (9), 280.

Andersen AN (1988) Dispersal distance as a benefit of myrmecochory. *Oecologia* **75**, 507–511.

Andersen AN (1995) A classification of Australian ant communities, based on functional groups which parallel plant life-forms in relation to stress and disturbance. *Journal of Biogeography* **22**, 15–29.

Barnes AD and Chapman HM (2014) Dispersal traits determine passive restoration trajectory of a Nigerian montane forest. *Acta Oecologica* **56**, 32–40.

Beck J and Chapman H (2008) A population estimate of the endangered chimpanzee Pan troglodytes vellerosus in a Nigerian montane forest: implications for conservation. *Oryx* **42**, 448–451.

Bernstein RA (1975) Foraging strategies of ants in response to variable food density. *Ecology* **56**, 213–219.

Bolton B (1994) *Identification Guide to the Ant Genera of the World*. Cambridge, MA: Harvard University Press.

Bowers MA and Dooley JL (1993) Predation hazard and seed removal by small mammals: microhabitat versus patch scale effects. *Oecologia* **94**, 247–254.

Bruhl CA, Mohamed M and Linsenmair KE (1999) Altitudinal distribution of leaf litter ants along a transect in primary forests on Mount Kinabalu, Sabah, Malaysia. *Journal of Tropical Ecology* **15**, 265–277.

Cain ML, Milligan BG and Strand AE (2000) Long-distance seed dispersal in plant populations. *American Journal of Botany* **87**, 1217–1227.

Camargo PH, Martins MM, Feitosa RM and Christianini AV (2016) Bird and ant synergy increases the seed dispersal effectiveness of an ornithochoric shrub. *Oecologia* **181**, 507–518.

Camargo PH, Rodrigues SB, Piratelli AJ, Oliveira PS and Christianini AV (2019) Interhabitat variation in diplochory: seed dispersal effectiveness by birds and ants differs between tropical forest and savanna. *Perspectives in Plant Ecology, Evolution and Systematics* **38**, 48–57.

Chapman H (2008) The Nigerian Montane Forest Project. *Tropinet* **19**, 7–9.

Chapman H, Cordeiro NJ, Dutton P, Wenny D, Kitamura S, Kaplin B and Lawes MJ (2016) Seed-dispersal ecology of tropical montane forests. *Journal of Tropical Ecology* **32**, 437–454.

Chapman JD, Chapman, JD and Chapman H (2001) *The forest flora of Taraba and Adamawa States, Nigeria: an ecological account and plant species checklist*. Department of Plant and Microbial Sciences, University of Canterbury.

Christian CE (2001) Consequences of a biological invasion reveal the importance of mutualism for plant communities. *Nature* **413**, 635–639.

Christianini AV and Oliveira PS (2010) Birds and ants provide complementary seed dispersal in a neotropical savanna. *Journal of Ecology* **98**, 573–582.

Christianini AV and Oliveira PS (2013) Edge effects decrease ant-derived benefits to seedlings in a neotropical savanna. *Arthropod–Plant Interactions* **7**, 191–199.

Christianini AV, Mayhé-Nunes AJ and Oliveira PS (2007) The role of ants in the removal of non-myrmecochorous diaspores and seed germination in a neotropical savanna. *Journal of Tropical Ecology* **23**, 343–351.

Christianini AV, Mayhé-Nunes AJ and Oliveira, PS (2012) Exploitation of fallen diaspores by ants: are there ant–plant partner choices? *Biotropica* **44**, 360–367.

Connell JH (1971) On the role of natural enemies in preventing competitive exclusion in some marine animals and in rain forest trees. *Dynamics of Populations* **298**, 312.

Cordeiro NJ and Howe HF (2003) Forest fragmentation severs mutualism between seed dispersers and an endemic African tree. *Proceedings of the National Academy of Sciences USA* **100**, 14052–14056.

Crawley MJ (2002) *Statistical Computing: An Introduction to Data Analysis using S-Plus* (No. 001.6424 C73).

Dunn RR, Agosti D, Andersen AN, Arnan X, Bruhl CA, Cerdá X ... Sanders NJ (2009) Climatic drivers of hemispheric asymmetry in global patterns of ant species richness. *Ecology letters* **12**, 324–333.

Farji-Brener AG and Werenkraut V (2017) The effects of ant nests on soil fertility and plant performance: a meta-analysis. *Journal of Animal Ecology* **86**, 866–877.

Gallegos SC, Hensen I and Schleuning M (2014) Secondary dispersal by ants promotes forest regeneration after deforestation. *Journal of Ecology* **102**, 659–666.

Giladi I (2006) Choosing benefits or partners: a review of the evidence for the evolution of myrmecochory. *Oikos* **112**, 481–492.

Gómez C and Espadaler X (2013) An update of the world survey of myrmecochorous dispersal distances. *Ecography* **36**, 1193–1201.

Gómez JM, Schupp EW and Jordano P (2019) Synzoochory: the ecological and evolutionary relevance of a dual interaction. *Biological Reviews* **94**, 874–902.

Gorb E And Gorb S (2003) *Seed Dispersal by Ants in a Deciduous Forest Ecosystem: Mechanisms, Strategies, Adaptations*. Dordrecht: Springer.

Gove AD, Majer JD and Dunn RR (2007) A keystone ant species promotes seed dispersal in a "diffuse" mutualism. *Oecologia* **153**, 687–697.

Guénard B, Weiser MD and Dunn RR (2012) Global models of ant diversity suggest regions where new discoveries are most likely are under disproportionate deforestation threat. *Proceedings of the National Academy of Sciences USA* **109**, 7368–7373.

Harrison XA (2014) Using observation-level random effects to model overdispersion in count data in ecology and evolution. *PeerJ* **2**, e616.

Hölldobler B and Wilson EO (1990) *The Ants*. Cambridge, MA: Harvard University Press.

Howe HF and Miriti MN (2000) No question: seed dispersal matters. *Trends in Ecology & Evolution* **15**, 434–436.

Hubbell SP, Foster RB, O'Brien ST, Harms KE, Condit R, Wechsler B ... De Lao SL (1999) Light-Gap disturbances, recruitment limitation, and tree diversity in a neotropical forest. *Science* **283**, 554–557.

Janzen DH (1970) Herbivores and the number of tree species in tropical forests. *American Naturalist* **104**, 501–528.

Kaspari M (1993) Body size and microclimate use in Neotropical granivorous ants. *Oecologia* **96**, 500–507.

Kenne M and Dejean A (1997) Caste polyethism and honeydew collection activity in foraging workers of *Myrmicaria opaciventris*. *Sociobiology* **30**, 247–255.

Downloaded from https://www.cambridge.org/core. University of Canterbury Library, on 29 Apr 2021 at 21:09:13, subject to the Cambridge Core terms of use, available at https://www.cambridge.org/core/terms. https://doi.org/10.1017/S0266467420000231

Kenne M and Dejean A (1999) Spatial distribution, size and density of nests of *Myrmicaria opaciventris* Emery (Formicidae, Myrmicinae). *Insectes Sociaux* **46**, 179–185.

Leal IR, Wirth R and Tabarelli M (2007) Seed dispersal by ants in the semi-arid Caatinga of north-east Brazil. *Annals of Botany* **99**, 885–894.

Leal LC, Andersen AN and Leal IR (2014) Anthropogenic disturbance reduces seed-dispersal services for myrmecochorous plants in the Brazilian Caatinga. *Oecologia* **174**, 173–181.

Levey DJ and Byrne MM (1993) Complex ant–plant interactions: rain-forest ants as secondary dispersers and post-dispersal seed predators. *Ecology* **74**, 1802–1812.

Magalhães VB, Espírito Santo NB, Salles LF, Soares Jr H and Oliveira PS (2018) Secondary seed dispersal by ants in Neotropical cerrado savanna: species-specific effects on seeds and seedlings of *Siparuna guianensis* (Siparunaceae). *Ecological Entomology* **43**, 665–674.

Mascaro J, Schnitzer SA and Carson WP (2004) Liana diversity, abundance, and mortality in a tropical wet forest in Costa Rica. *Forest Ecology and Management* **190**, 3–14.

Matthesius A (2006) Testing the Janzen–Connell model for species diversity in a West African montane forest. MSc thesis, University of Canterbury, New Zealand.

Matthesius A, Chapman H and Kelly D (2011) Testing for Janzen–Connell effects in a west African montane forest. *Biotropica* **43**, 77–83.

Moles AT and Westoby M (2004) What do seedlings die from and what are the implications for evolution of seed size? *Oikos* **106**, 193–199.

Nathan R and Muller-Landau HC (2000) Spatial patterns of seed dispersal, their determinants and consequences for recruitment. *Trends in Ecology & Evolution* **15**, 278–285.

Ness JH, Bronstein JL, Andersen AN and Holland JN (2004). Ant body size predicts dispersal distance of ant-adapted seeds: implications of small-ant invasions. *Ecology* **85**, 1244–1250.

O'Dowd DJ and Gill AM (1984) Predator satiation and site alteration following fire: mass reproduction of alpine ash (*Eucalyptus delegatensis*) in southeastern Australia. *Ecology* **65**, 1052–1066.

Oliveira PS, Galetti M, Pedroni F and Morellato LPC (1995) Seed cleaning by *Mycocepurus goeldii* ants (Attini) facilitates germination in *Hymenaea courbaril* (Caesalpiniaceae). *Biotropica* **27**, 518–522.

Orrock JL and Christopher CC (2010) Density of intraspecific competitors determines the occurrence and benefits of accelerated germination. *American Journal of Botany* **97**, 694–699.

Palfi Z, Spooner PG and Robinson W (2017) Seed dispersal distances by ants increase in response to anthropogenic disturbances in Australian roadside environments. *Frontiers in Ecology and Evolution* **5**, 132.

Passos L and Oliveira PS (2002) Ants affect the distribution and performance of seedlings of *Clusia criuva*, a primarily bird-dispersed rain forest tree. *Journal of Ecology* **90**, 517–528.

Passos L and Oliveira PS (2003) Interactions between ants, fruits and seeds in a restinga forest in south-eastern Brazil. *Journal of Tropical Ecology* **19**, 261–270.

Passos L and Oliveira PS (2004) Interaction between ants and fruits of *Guapira opposita* (Nyctaginaceae) in a Brazilian sandy plain rainforest: ant effects on seeds and seedlings. *Oecologia* **139**, 376–382.

Pirk GI, Di Pasquo F and de Casenave JL (2009) Diet of two sympatric *Pheidole* spp. ants in the central Monte desert: implications for seed–granivore interactions. *Insectes Sociaux* **56**, 277.

Pizo MA and Oliveira PS (1998) Interaction between ants and seeds of a non-myrmecochorous neotropical tree, *Cabralea canjerana* (Meliaceae), in the Atlantic forest of southeast Brazil. *American Journal of Botany* **85**, 669–674.

Pizo MA and Oliveira PS (2001) Size and lipid content of nonmyrmecochorous diaspores: effects on the interaction with litter-foraging ants in the Atlantic rain forest of Brazil. *Plant Ecology* **157**, 37–52.

Raimundo RLG, Guimaraes PR, Almeida-Neto M and Pizo MA (2004) The influence of fruit morphology and habitat structure on ant-seed interactions: a study with artificial fruits. *Sociobiology* **44**, 261–270.

Roberts JT and Heithaus ER (1986) Ants rearrange the vertebrate-generated seed shadow of a neotropical fig tree. *Ecology* **67**, 1046–1051.

Sabu TK, Shiju, RT, Vinod KV and Nithya S (2011) A comparison of the pit-fall trap, Winkler extractor and Berlese funnel for sampling ground-dwelling arthropods in tropical montane cloud forests. *Journal of Insect Science* **11**. doi: 10.1673/031.011.0128.

Seidler TG and Plotkin JB (2006) Seed dispersal and spatial pattern in tropical trees. *PLoS Biology* **4**, e344.

Smith DS, Lau MK, Jacobs R, Monroy JA, Shuster SM and Whitham TG (2015) Rapid plant evolution in the presence of an introduced species alters community composition. *Oecologia* **179**, 563–572.

Svenning JC, Normand S and Skov F (2006) Range filling in European trees. *Journal of Biogeography* **33**, 2018–2021.

Symonds MR and Moussalli A (2011) A brief guide to model selection, multimodel inference and model averaging in behavioural ecology using Akaike's information criterion. *Behavioral Ecology and Sociobiology* **65**, 13–21.

Tamme R, Götzenberger L, Zobel M, Bullock JM, Hooftman DA, Kaasik A and Pärtel M (2014) Predicting species' maximum dispersal distances from simple plant traits. *Ecology* **95**, 505–513.

Thomson FJ, Auld TD, Ramp D and Kingsford RT (2016) A switch in keystone seed-dispersing ant genera between two elevations for a myrmecochorous plant, *Acacia terminalis*. *PLoS ONE* **11**, e0157632.

Van Schaik CP, Terborgh JW and Wright SJ (1993) The phenology of tropical forests: adaptive significance and consequences for primary consumers. *Annual Review of Ecology and Systematics* **24**, 353–377.

Verdú M and Traveset A (2005) Early emergence enhances plant fitness: a phylogenetically controlled meta-analysis. *Ecology* **86**, 1385–1394.

Vittoz P and Engler R (2007) Seed dispersal distances: a typology based on dispersal modes and plant traits. *Botanica Helvetica* **117**, 109–124.

Wang BC and Smith TB (2002) Closing the seed dispersal loop. *Trends in Ecology & Evolution* **17**, 379–386.

Warren RJ and Giladi I (2014) Ant-mediated seed dispersal: a few ant species (Hymenoptera: Formicidae) benefit many plants. *Myrmecological News* **20**, 129–140.

Weckerle CS and Rutishauser R (2005) Gynoecium, fruit and seed structure of Paullinieae (Sapindaceae). *Botanical Journal of the Linnean Society* **147**, 159–189.

Wenny DG (2001) Advantages of seed dispersal: a re-evaluation of directed dispersal. *Evolutionary Ecology Research* **3**, 37–50.

Zamble A, Carpentier M, Kandoussi A, Sahpaz S, Petrault O, Ouk T and Martin-Nizard F (2006) *Paullinia pinnata* extracts rich in polyphenols promote vascular relaxation via endothelium-dependent mechanisms. *Journal of Cardiovascular Pharmacology* **47**, 599–608.

Zelikova TJ (2008) Variation in the effects of ants on seed dispersal and ecosystem processes. Thesis. University of Colorado at Boulder.

Zelikova TJ and Breed MD (2008) Effects of habitat disturbance on ant community composition and seed dispersal by ants in a tropical dry forest in Costa Rica. *Journal of Tropical Ecology* **24**, 309–316.

Zwiener VP, Bihn JH and Marques M (2012) Ant–diaspore interactions during secondary succession in the Atlantic forest of Brazil. *Revista de Biologia Tropical* **60**, 933–942.

Downloaded from https://www.cambridge.org/core. University of Canterbury Library, on 29 Apr 2021 at 21:09:13, subject to the Cambridge Core terms of use, available at https://www.cambridge.org/core/terms. https://doi.org/10.1017/S0266467420000231